## FISHERIES RESEARCH INSTITUTE

SCHOOL OF FISHERIES WH-10
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## ALASKA SALMON RESEARCH

D. ROGERS, PRINCIPAL INVESTIGATOR; T. QUINN, ASSOCIATE PROFESSOR; C. FOOTE, ASSISTANT PROFESSOR; AND B. ROGERS , FISHERY BIOLOGIST

ANNUAL REPORT-1993
to
PACIFIC SEAFOOD PROCESSORS ASSOCIATION

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## CONTENTS

Page
LIST OF FIGURES ..... iv
LIST OF TABLES ..... v
KEY WORDS ..... vi
INTRODUCTION ..... 1
FORECASTING .....  1
Pre-season Forecasts ..... 1
LAKE RESEARCH ..... 3
Kvichak System ..... 3
Newhalen River Escapement. ..... 3
Spawning Ground Surveys ..... 4
Sockeye Fry Abundance and Size. ..... 5
Sculpin Predation ..... 5
Spawning Behavior, Morphology and Genetics ..... 6
Wood River System ..... 7
Environmental Observations ..... 7
Fry Abundance and Growth ..... 7
Char Predation ..... 8
Spawning Ground Surveys ..... 8
Bear Predation ..... 8
LITERATURE CITED ..... 10

## LIST OF FIGURES

Page Figure

1. Reconstructed daily Bristol Bay sockeye runs at Port Moller and the daily sockeye salmon index catches, 1990-1993 ..... 13
2. Timing of the sockeye salmon escapements to Wood River, 1908-1919 and 1953-1993 ..... 14
3. Annual escapements of sockeye salmon to the Wood River Lakes and annual stream survey counts for all Lake Aleknagik creeks and Hansen Creek, 1949-1993 ..... 15

## LIST OF TABLES

Table Page

1. Pre-season forecasts of Bristol Bay sockeye salmon inshore runs ..... 19
2. Bristol Bay sockeye salmon runs and the predictions from the Port
Moller test-boat catches ..... 20
3. Mean lengths of sockeye salmon in the Bristol Bay area ..... 21
4. Average weights of sockeye in the east-side Bristol Bay commercial catches, 1984-1993 ..... 22
5. Average weights of sockeye in the west-side Bristol Bay commercial catches, 1984-1993 ..... 23
6. Sockeye salmon escapements in excess of management goals for Bristol Bay river, 1987-1993 ..... 24
7. Cumulative daily escapements of sockeye salmon in the Kvichak and Newhalen Rivers, 1989-1993 ..... 25
8. The Kvichak lake system escapements and the percentages going to the Newhalen River and Lake Clark ..... 26
9. Age compositions of sockeye salmon on the Kvichak spawning grounds in 1993 ..... 27
10. Spawning ground estimates of sockeye salmon on 29 selected spawning grounds in Lake Iliamna and the Newhalen River system, 1956-1993 ..... 28
11. Mean townet catches and lengths on Sept. 1 of sockeye salmon fry in lakes Iliamna and Clark ..... 29
12. Summary of 1993 measurements in Lake Aleknagik ..... 30
13. Five-day averages of catches of emergent midges and water temperatures at three stations on Lake Aleknagik, 1993 ..... 31
14. Average catches, lengths, and growth rates for sockeye fry and age-1 threespine stickleback in Lake Aleknagik ..... 32
15. Average townet catches and mean lengths of sockeye fry, numbers of parent spawners, and average catches and mean lengths of threespine stickleback for Lake Nerka ..... 33
Table ..... Page
16. Occurrence and numbers of juvenile sockeye in stomachs of Arctic char collected by hook and line from Little Togiak River during 30 days after ice-out ..... 34
17. Ground survey counts of sockeye spawners in the Wood River Lakes, 1993 ..... 35
18. Daily counts of sockeye salmon spawners in Hansen Creek, 1993 ..... 36
19. Summary of Hansen Creek spawning surveys, 1990-93 ..... 37

## KEY WORDS

Bristol Bay, Kvichak, Nushagak, escapements, growth, predation, sockeye salmon, forecasts, spawning, genetics, Wood River Lakes

## ALASKA SALMON RESEARCH

## INTRODUCTION

Fisheries Research Institute was established in 1946 with the financial support of the major Alaskan salmon processors to (1) investigate the causes of the declines in production that had occurred in most stocks since the 1930s, (2) work with the government management agency to increase our knowledge of the biology of salmon and the effects of the fisheries on the stocks, and (3) assist salmon processors by providing a second opinion on matters of salmon fisheries management. With the high levels of production since the 1980s, our primary objectives now are to determine how to maintain the high production (what has caused year-to-year variation) and how to harvest/process salmon most efficiently, e.g., accurate forecasts and fishing evenly distributed throughout the run.

We presently have salmon research projects in Bristol Bay, the Alaska Peninsula, Southeast Alaska and Chignik that are funded in part or entirely by the industry. In addition, we have a federally funded high seas salmon project that is concerned with the oceanic distribution of salmon and the vulnerability of North American stocks to foreign fisheries. In recent years, we have also worked at Kodiak and on the Yukon stocks. All of these projects have been carried out in cooperation with the Alaska Department of Fish and Game (ADF\&G) or the National Marine Fisheries Service (High Seas), and we have also had cooperative research projects with salmon biologists from Japan and Russia.

This report will focus on our 1993 Bristol Bay research with emphasis on salmon forecasting and research relevant to escapement policies to maximize production. The Southeast pink salmon (Oncorhynchus gorbuscha) research will be reported in a separate report from the University of Alaska; our Chignik salmon research is reported to the National Marine Fisheries Service; and a report on our Alaska Peninsula (False Pass) work we be completed soon.

## FORECASTING

## PRE-SEASON FORECASTS

Forecasts of the 1993 Bristol Bay sockeye salmon (O. nerka) runs and catches were provided to participating processors at our October 1992 meeting. They are presented in Table 1 with the ADF\&G forecasts and the past forecasts and runs beginning in 1986. The two river system forecasts (FRI and ADF\&G) are based on the same data sources, but different analytical methods have often been used. Both 1993 forecasts were for an average (recent years) run and catch, whereas the actual run was above average and the catch ( 41 million) was the largest on record. Since 1986, the actual catch has been between the two forecasted catches only once. In the other 7 years, the catch was either higher (6) or lower (1) than both forecasts. The 1993 catch was $28 \%$ greater than our forecast, which was near the average error since 1986 (31\%). Our forecasts of the
runs to the fishing districts have been closest for the Nushagak and farthest off for the Naknek/Kvichak, with the error in Egegik and Ugashik forecasts being intermediate.

## PORT MOLLER FORECAST

The Port Moller in-season test fishery was conducted by ADF\&G during June and early July from 1968 through 1985, with a change in gear in 1985. There was no test fishery in 1986 and, beginning in 1987, we have conducted the test fishery each year. The accuracy of the forecasts since 1987 has been very good. The runs have differed from the forecasts made on June 25 and 30 by an average of $16 \%$, and we have been within an average of $10 \%$ on forecasts made about July 3 (Table 2). We have not done as well in forecasting the catch because river system forecasts and thus catches cannot be made until about July 3, when we have the first indication of where the salmon are going. In 1993, the test-fish catches projected that the run would be earlier and larger than the pre-season forecasts. Daily forecasts, from an almanac provided to Bristol Bay processors, tended to over-forecast the run until July 6 when the run was forecast exactly ( 52 million). Periodic adjusted forecasts (to account for the early run timing) were lower than the actual run by about $15 \%$.

The test fishery at Port Moller employs a 200-fathom gillnet that is 60 meshes deep and has 5-in stretched mesh. The web is multistrand monofilament (center core). We have used a $70-\mathrm{ft}$ vessel (Nettie H) and fished each day from June 11 through about July 5 (weather permitting). Four stations are fished along a transect $\sim 30-60 \mathrm{mi}$ out from Port Moller. Catch, mean length, and water temperature data are sent daily by radio to Port Moller and then faxed into Bristol Bay. The vessel comes into port every other day to deliver fish and salmon scales collected by the 2-3 biologists onboard. In 1993, scales and length data were sent periodically to ADF\&G (B. Cross, King Salmon), and the scales were aged and the age compositions and average lengths by age were reported.

The statistics from Port Moller in 1993 were again a challenge to interpret, although they ultimately predicted the total run and age composition accurately. The sockeye salmon were larger than average for their age, and there was a higher than expected percentage of age 2.3 fish. Usually high percentages of 3-ocean fish and large body size are associated with small runs, but there were large index catches (especially early, June 14-17) indicating that a large run was coming with an early arrival in the bay. ADF\&G provided preliminary daily catches and escapements for 1993, and from these data as well as published statistics (e.g., Stratton 1991) we reconstructed the run timing in the Bristol Bay fishing districts to compare with past years and with the Port Moller index catches (Fig. 1). The timing and magnitude of the 1993 run was fairly well predicted by the Port Moller catches assuming a 6-day travel time and that Ugashik fish passed Port Moller at the same time as the other stocks (Rogers 1990). In 1993, the fish arrived early off Port Moller and then took about 1 or 2 days less than average to reach the fishing districts.

ADF\&G (B. Stratton, Anchorage) provided preliminary length and weight statistics for 1993, and statistics from prior years were available (e.g., Yuen et al. 1981 and Stratton 1991) so that we could calculate mean lengths in the runs (Table 3). The 2-ocean sockeye salmon in the 1993 run were much larger than average, whereas the 3 -ocean fish were a little below average in length. Large runs typically contain smaller fish because of density-dependent growth in their final spring
at sea (Rogers 1980). Because there were high percentages of 3-ocean fish in the catches, the average weights calculated from ADF\&G sampling were above average except in the Nushagak District (Tables 4 and 5).

The Port Moller test fishery in 1993 provided an early indication to ADF\&G management that an early and large run was on the way; however, the early distribution of the run, heavy to the Egegik District, created some doubt as to the size of the run (40-50 million). Although there was some overescapement in the Naknek, Egegik, Ugashik, and Igushik rivers, the escapements in 1993 were closer to the goals than in any of the past 4 years (Table 6). Considering the size of the run (a daily record of 5.3 million landed on July 2 ) and the early timing (second to 1979), management of the run was outstanding. Good catches were made in all districts before large numbers of fish were counted past the towers. The timings of most escapements were the earliest recorded. For Wood River, the 1993 run continued a trend towards earlier timing (Fig. 2). Escapements during the early 1900s were $\sim 1$ wk later than recent years (since 1953), and the escapements since 1977 have begun early ( $10 \%$ date) except for 1986 . Two factors contribute to the recent early timing: (1) a majority of the years have had early or warm spring weather, and (2) there is a tendency to keep fishing closed until a significant number of fish have been counted past the upriver tower sites. The later timing of the escapements in the early 1900 s was partly caused by the reverse situation (i.e., the fishery tended to fish more on the early part of the run than on the late part).

## LAKE RESEARCH

During the summer of 1993, we continued our long-term studies of spawner distribution, growth and abundance of fry, and the physical and biological environment for the sockeye salmon of the Wood River (Nushagak) and Kvichak lake systems. Most of our annual observations in the Wood River Lakes extend over more than 30 years and constitute the longest continuous biological and environmental record on any salmon stock in Alaska. In 1993, we also conducted special studies of bear predation on spawning sockeye salmon and stock specific traits of sockeye spawning populations.

## KVICHAK SYSTEM

Our 1993 field season in the Kvichak system (Lakes Iliamna and Clark) consisted of estimating the sockeye escapement into the Newhalen River in late June and July, townetting for juvenile sockeye and threespine stickleback (Gasterosteus aculeatus) in upper Lake lliamna and for sockeye fry, stickleback and least cisco (Coregonus sardinella) in Lake Clark in August. We also conducted spawning ground surveys in late August-early September to collect otoliths for age determination. We continued our studies on (1) the ecological relationship between sockeye salmon and two sculpin species, Cottus cognatus and C. aleuticus, and (2) factors promoting the genetic differentiation of sockeye salmon populations.

## Newhalen River Escapement

The annual escapements of sockeye salmon to the Kvichak lake system are estimated by ADF\&G from expanded $10-\mathrm{min}$ counts on each bank of the river at the outlet of Lake Iliamna (Igiugig). In addition, since 1979 we have estimated the escapements up the Newhalen River by expanding 20-
min counts on one bank, for each of 10 daylight hours, to a daily count for both banks. We count when and where the visibility is best and assume that the fish utilize both banks equally and that their migratory rate does not change at night. The daily counts at Newhalen are compared with the counts at Igiugig to estimate a travel time; then, by lagging the Newhalen counts back to Igiugig the appropriate number of days, we can calculate the daily proportions of the Kvichak run that went up the Newhalen River.

The cumulative daily escapements for the two rivers, timed to the Kvichak, are given in Table 7 for 1989-1993. In mid-July, milling fish often swim upriver along the banks of the Newhalen and are counted, and then drift down river in the middle where they cannot be seen, only to swim up river again. This inflates the counts for the escapement; therefore, we have used the average proportion of Newhalen count to Kvichak count for day 5 to day 16 (day 1 equals the first day of 100,000 in the Kvichak) and the season's total Kvichak escapement to estimate the Newhalen/Lake Clark escapement.

In 1993, we estimated that 1.6 million of the Kvichak escapement of 4.0 million (about $38 \%$ ) ended up in the Newhalen/Lake Clark system (Table 8). This was about average for the past 5 yr (35\%). The aerial surveys conducted by ADF\&G in 1993 provided an estimate of the Newhalen River spawners and, thus by subtraction, an estimate of the Lake Clark escapement of 1.5 million.

## Spawning Ground Surveys

We have collected scales or otoliths from spawned out sockeye salmon from several major spawning grounds in the Kvichak system each year since 1956. In 1993, 8 spawning grounds were sampled and the age compositions from the samples provided a similar pattern to the age composition in the lake system escapement (Table 9). Chinkelyes Creek had a high percentage of age 2.2, the Newhalen and Tazimina Rivers had high percentages of age 1.3, and Pedro Creek had a high percentage of age 1.2; otherwise the age compositions were similar to the composition for the entire lake system (Kvichak escapement).

We had conducted annual aerial surveys of the Kvichak spawning grounds from 1956 until 1988, when ADF\&G took over the surveys. The results of the 1993 surveys were provided by ADF\&G (J. Regnart, King Salmon) and are summarized for 29 selected spawning grounds in Table 10. In recent years the surveys have accounted for smaller percentages of the total (tower count) escapement than was typical for past years. This may have been caused by differences in observers, weather conditions (visibility), or distribution of sockeye salmon on the spawning grounds.

We continued our survey of the spawning dynamics of sockeye salmon on the beaches of Fuel Dump Island. Preliminary results of 1993 indicated the possibility of local differentiation in spawning time between groups of sockeye salmon that were within 20 m of each another. We repeated our detailed transect counts of the area by snorkeling and collected various sets of physical data (gravel size, water temperature and measures of current). In addition, we tagged males within the two areas to measure the intermixing between groups and hence the possibility for stock differentiation. As in 1992, there was a difference in spawning time of a few days between two adjacent sites. Physical factors also differed between areas. In particular, water temperature commonly differed by nearly a degree (C) and the current flows differed considerably.

## Sockeye Fry Abundance and Size

We have sampled the sockeye fry (age 0) in the Kvichak system each year since 1962 (1961 brood year) by townetting at night. We towed in both Lake Iliamna and Lake Clark, but as usual we did not sample the fry in Six-mile Lake (upper end of the Newhalen River), where fry from the Tazimina River are likely to concentrate. The geometric means of the catches provide a measure of the relative density (number per 20-min. tow), and the mean lengths of the fry are adjusted each year, based on their daily growth rate, to September 1 (Table 11).

The sockeye fry are usually smaller in Lake Clark than in Lake Iliamna because temperatures are usually colder and Lake Clark has a shorter ice-free period; however this was not so in 1993 (1992 brood year). The fry in Lake Clark averaged a record 61 mm . In both lakes, the annual growth of the fry is correlated with water temperatures, which are mostly influenced by spring weather. Cold temperatures typically result in small fry ( $40-50 \mathrm{~mm}$ ), which then spend 2 yr in the lake before seaward migration and tend to return as adults 5 yr after their parents. Warm temperatures usually result in large fry ( $>60 \mathrm{~mm}$ ), which tend to migrate to sea after 1 yr and mostly return 4 yr after their parents. The townet sampling has been useful in predicting, 3 yr in advance, the main age at return from the larger Kvichak escapements by utilizing the relationship between age at return and mean length of fry in Lake Iliamna. From the mean lengths in 1993 ( 57 and 61 mm ), we would expect most of the fish from the 1992 brood year to migrate as age 1 smolt and return mainly in 1996.

## Sculpin Predation

In 1992 we examined the movement patterns and distribution of sculpins relative to sockeye spawning grounds, while concurrently measuring the single meal egg feeding potential of sculpins in laboratory conditions. The results indicated that sculpins could have a major impact on sockeye egg-to-fry survival on these beaches. In 1993 we extended the study four ways. First we estimated the maximum number of eggs sculpins could consume over the course of the spawning season. Second, we measured the digestion rate of both species when fed fresh and waterhardened eggs. Third, we estimated indirectly the hunger level of sculpins on the spawning grounds over the course of the spawning period, and hence the susceptibility of eggs to predation. Finally, we estimated the density of sculpins in relation to spawning site characteristics (gravel size) to determine if the intensity of egg predation varies among sites (habitats).

The number of eggs consumed by sculpins over the course of the spawning season increased in relation to sculpin size, with the largest sculpins ( 11 cm ) consuming over 115 eggs in a 2 -wk period. However, egg digestion was relatively slow, with residue of meals of five eggs sometimes still evident 6 d after feeding. The tendency of sculpins to fill up rapidly and digest slowly probably accounts for the marked decrease in the susceptibility of eggs to predation over the course of the spawning period. While 40 g of eggs in minnow traps attracted on average over 150 sculpins per trap when set overnight just before spawning started, the same traps often attracted no sculpins 5 d after spawning had begun. This suggests that sculpins were satiated once spawning was underway. However, the number of sculpins increased in the traps near the end of the spawning period, indicating that hunger levels were increasing.

This study indicates that sculpin predation can have a significant impact on egg to fry survival. Furthermore, the intensity of this egg predation is not random across the spawning period; fish spawning at the beginning and end of the spawning period undoubtedly suffer higher egg predation than those spawning in the middle. Hence, natural selection, in the form of sculpin predation, will favor sockeye salmon that spawn over a narrow period of time. This may account for the short spawning periods observed on island beaches where sculpins are abundant compared with the long periods of spawning observed in Knutson Bay and various rivers where sculpin densities are low. In the future, we will attempt to asses sculpin densities through depletion estimates. This will also test the feasibility of reducing local sculpin densities, thus reducing egg predation and increasing egg to fry survival.

## Spawning Behavior, Morphology and Genetics

In 1993 we continued our work on sockeye salmon spawning behavior on island beaches, explored additional hypotheses to account for population differentiation in spawning morphology, initiated studies on physical factors affecting egg to fry survival and completed our preliminary examination of genetic variation (DNA) among sockeye salmon spawning populations in Bristol Bay.

We completed our work examining the behavior of males relative to females over the course of the spawning period. Unlike river spawning fish, those on the beaches do not adopt sneak positions relative to females when denied access by other males. Rather, as the number of males available per female increased over the spawning period, large aggregations of males (often over 10) began to appear. There were no obvious dominant males within these groups; rather all males attempted to participate in lone female spawnings. This caused repeated displacement of the female and a protracted individual spawning period.

Since 1988 we have explored hypotheses regarding population differentiation in spawning morphology. We have examined the significance of this variation within populations and described the variation among populations and related it to spawning habitat. In 1993 we examined the effect of parasite infestation on body shape for spawning populations in Lake Iliamna and Aleknagik (Wood River Lakes). Parasite infestation of Philonema oncorhynchi was directly related to freshwater age. Adults that had spent 2 yr in the lake prior to seaward migration had higher rates of infestation than those that spent only 1 yr in the lake. However, infestation rate was not related to body shape, indicating that the parasite caused a negligible cost to the host (sockeye salmon adult). This provided additional support that differences among populations in body morphology are likely genetic in origin and reflect adaptations to local spawning sites.

Preliminary DNA analyses of populations from several areas indicated that sockeye salmon populations from Lake Clark, Lake Iliamna, and Lake Aleknagik were genetically distinct from one another. Further, these Bristol Bay populations were more closely related to sockeye salmon from Russia than populations from southerly locations in North America. This technique (DNA analysis) holds great promise for identifying Bristol Bay stocks within the fisheries or at Port Moller.

In future work, we will examine physical and biotic factors that affect egg-to-fry survival (gravel size, flow and egg size). We will also examine the long-term effects of differentiation in egg size
among populations by examining genetic differentiation in development rates in controlled experiments at the University of Washington, and by measuring individual growth rates of fry in the wild (daily growth rings on otoliths) in relation to their parental spawning location.

## WOOD RIVER SYSTEM

The Bristol Bay research program of FRI began with spawning ground surveys in the Wood River Lakes in 1946 to determine where, when, and how many sockeye spawned there. During the early 1950s, methods were established to enumerate and sample the commercial catches, escapements (towers), and the smolts produced. By the late 1950s, we had established several important measurements, which we have maintained to the present in order to characterize each year's environment for spawning adults and rearing juveniles.

## Environmental Observations

The spring of 1993 was the earliest for Bristol Bay since 1981. The April air temperature was the second warmest in history (since 1919) and ice breakup in Lake Aleknagik (recorded since 1949) was 1 mo earlier than average and 2 wk earlier than the previous record of May 14, 1958 (Table 12). Although ice breakup was early, mid-summer water temperatures were only a little warmer than average because solar radiation (sunlight) was below average during most of the summer. Lake levels were about normal until late August, when heavy rainfall caused a moderate increase. Standing crop of phytoplankton (chlorophyll) was about average during the summer, whereas zooplankton volumes were well above average in June and July but below average in August. Zooplankton are the main source of food for juvenile sockeye salmon after they move offshore in late July. Insects (mainly pupal and adult midges) are the main source of food in the spring, when the fry are inshore. There was an early peak in midge emergence in 1993 (early July) corresponding to the early ice breakup. In past years midge emergence has usually peaked in either late July or early August (Table 13). Water temperature at nearshore insect traps were warmer than average in 1993, but no records were set.

## Fry Abundance and Growth

The sockeye fry in Lake Aleknagik in 1993 were longer than average in June, but their growth during July and August was only average, and on September 1 their lengths were typical of past years (Table 14). Fry abundance as measured by beach seine and townet sampling was below average, although the number of parent spawners $(343,000)$ in Lake Aleknagik in 1992 was above average for the lake. The relatively small size on September 1 indicates that the fry and sticklebacks had cropped down their main food supply, especially the larger forms of the zooplankton such as calanoid copepods and Daphnia .

The mean lengths of sockeye salmon fry in Lake Nerka indicated that in 1993 growth was about average, but the relative abundance of fry was below average as estimated from townet catches (Table 15). Juvenile sockeye salmon in the Wood River Lakes system exhibit density-dependent growth, and we are analyzing our long-term data set for Lake Aleknagik to determine the relative effects of physical and biological factors in the lake on the growth of the sockeye salmon fry. In addition, we are comparing year-to-year variation in zooplankton population composition (1993 samples have not yet been counted) with annual variation in sockeye salmon fry and threespine stickleback abundance to determine the extent to which the fish alter their food resources. We hope
the information from these studies will help explain the observed variability in the freshwater phase of the sockeye salmon.

## Char Predation

We have sampled the Arctic char (Salvelinus alpinus) in Little Togiak River each spring since 1972 to follow the rate of predation on juvenile sockeye, especially smolts. This short river flows from Little Togiak Lake into Lake Nerka, and the smolts are very vulnerable to the char for the few minutes it takes them to move from one lake to the next. Large char usually eat more juvenile sockeye than small char. The char caught in 1993 were about average in length and they consumed about 1 smolt per d (Table 16). We were surprised to find sockeye salmon fry in the char stomachs in as late as mid-June in 1993 because ice breakup was so early. Perhaps timing of fry emergence in the river was normal in 1993. There are about 5,000 char in and around the river mouth, so that at just 1 smolt per char per night for a migration of 20 to 30 d , a significant number of smolts are lost from the production of this small lake in the system.

## Spawning Ground Surveys

Sockeye salmon spawning ground surveys have been conducted annually in the Wood River Lakes system since 1946; however it was not until the early 1950 s that all of the major spawning grounds were included. We collect otoliths from the major spawning grounds for age determination and make ground counts of the number of spawners in the small streams. ADF\&G estimates the numbers of spawners on the lake beaches and in the interconnecting rivers by aerial surveys; thus the total escapement to the lake system can be apportioned to the individual lakes or type of spawning ground (creek, river, and beach). The distribution of spawners among the lakes is used in forecasting the Wood River runs. Even escapement distributions tend to produce large returns and uneven distributions tend to produce small returns.

Aerial surveys were conducted by ADF\&G in 1993 with industry funding; however, we have not yet compiled the aerial survey counts to estimate the escapements to the lakes. The ground survey counts for the major spawning grounds in 1993 are given in Table 17. The creeks draining into Lake Aleknagik again contained relatively high counts of spawners (especially Happy Creek) and Hansen Creek had a large number of spawners for the fourth consecutive year (Fig. 3).

## Bear Predation

We completed the fourth year of our bear/spawning sockeye salmon interaction study in Hansen Creek, a small tributary of Lake Aleknagik where predation by bears is high relative to larger creeks. During 18 July to 22 August, a large number of spawners were again observed in Hansen Creek, but not as many in the prior 3 yr (Table 18). Daily count and removal of sockeye salmon killed by bears indicated that $1,504(36 \%)$ of 4,212 spawners were killed by bears in 1993 (Table 19). These estimates excluded dead fish from previous daily surveys that might have been attacked by bears (decisions to exclude fish were based on gill and body coloration, body firmness, and body deterioration). Numbers of sockeye killed by bears in 1993 was similar to the numbers in 1990 or 1992; however, the percentage of sockeye salmon killed by bears in 1993 was higher than in the past $3 \mathrm{yr}(16 \%$ to $24 \%)$. We plan to continue the daily surveys in Hansen Creek until we obtain counts for a year when number of spawners is near the median number $(2,500)$ and a year when there is a small number of spawners $(<1,000)$.

The daily counts on Hansen Creek are also providing us with estimates of the percentages of the total number of spawners that are counted on a single "peak survey" date and thus a means of adjusting our annual survey counts to equal the true number of spawners. Hansen Creek has been surveyed most often on August 6 in past years; but in $20 \%$ of the years, the survey was done on August 1 or earlier. The Hansen Creek sockeye are about the earliest spawners in the lake system and the fish usually first enter the creek about July 22-25. From the daily counts in 1990-1993, had the surveys been conducted on the single date of August 6 , the "peak survey" counts would have been $72 \%$ to $89 \%$ of the totals; if the single surveys were done on August 1, the counts would have been $38 \%$ to $78 \%$ of the actual number of spawners (Table 19). The percentage counted in 1992 on August $1(78 \%)$ was relatively high because spawning was early with the fish first entering the creek on July 18.

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## FIGURES



Figure 1. Reconstructed daily Bristol Bay sockeye runs at Port Moller and the daily sockeye salmon index catches (scaled: 1 index $=15,000$ ), 1990-1993.


Figure 2. Timing of the sockeye salmon escapements to Wood River, 1908-1919 and 19531993.


Figure 3. Annual escapements of sockeye salmon to the Wood River Lakes and annual stream survey counts for all Lake Aleknagik creeks and Hansen Creek, 1949-1993.

TABLES

Table 1. Pre-season forecasts of Bristol Bay sockeye salmon inshore runs (millions).

| Year | Forecast/run | Kvichak | Naknek | Egegik | Ugashik | Nushagak | Total run | Catch | \% error |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1985 | FRI | 12.2 | 5.3 | 5.8 | 4.4 | 4.3 | 33.0 | 18.2 | 29 |
|  | ADFG | 12.2 | 4.9 | 6.6 | 5.6 | 4.3 | 35.0 | 20.3 |  |
|  | Actual run | 13.4 | 3.7 | 8.6 | 7.4 | 3.0 | 36.6 | 23.5 |  |
| 1986 | FRI | 9.2 | 4.5 | 5.9 | 6.7 | 4.8 | 32.1 | 19.4 | 19 |
|  | ADFG | 4.5 | 3.2 | 5.4 | 4.9 | 3.8 | 22.5 | 13.3 |  |
|  | Actual run | 2.0 | 3.9 | 6.0 | 6.0 | 4.8 | 23.5 | 15.7 |  |
| 1987 | FRI | 2.8 | 2.0 | 5.8 | 3.1 | 5.1 | 19.5 | 12.4 | 29 |
|  | ADFG | 2.7 | 2.1 | 4.9 | 3.1 | 3.3 | 16.8 | 9.3 |  |
|  | Actual run | 9.6 | 2.4 | 6.6 | 2.8 | 5.2 | 27.4 | 16.0 |  |
| 1988 | FRI | 12.3 | 3.1 | 6.2 | 3.1 | 5.0 | 30.6 | 20.8 | 34 |
|  | ADFG | 9.3 | 2.5 | 5.6 | 3.2 | 5.6 | 26.5 | 16.8 |  |
|  | Actual run | 6.7 | 1.7 | 8.1 | 2.2 | 3.2 | 23.0 | 13.8 |  |
| 1989 | FRI | 20.4 | 3.6 | 6.7 | 3.0 | 3.4 | 38.0 | 25.4 | 13 |
|  | ADFG | 12.5 | 3.1 | 5.6 | 3.6 | 3.1 | 28.9 | 16.2 |  |
|  | Actual run | 19.8 | 3.2 | 10.5 | 4.9 | 5.0 | 43.9 | 28.7 |  |
| 1990 | FRI | 10.1 | 4.8 | 6.6 | 3.0 | 4.6 | 29.8 | 19.0 | 74 |
|  | ADFG | 8.9 | 3.6 | 5.6 | 3.1 | 3.5 | 25.4 | 14.7 |  |
|  | Actual run | 17.4 | 8.4 | 12.3 | 2.9 | 5.7 | 47.6 | 33.1 |  |
| 1991 | FRI | 12.0 | 4.6 | 8.9 | 3.6 | 6.9 | 36.7 | 25.0 | 5 |
|  | ADFG | 7.6 | 6.0 | 8.2 | 3.5 | 3.8 | 30.0 | 21.2 |  |
|  | Actual run | 8.1 | 10.0 | 9.6 | 5.5 | 7.7 | 42.1 | 26.2 |  |
| 1992 | FRI | 10.2 | 3.2 | 10.4 | 4.0 | 4.3 | 33.0 | 22.0 | 45 |
|  | ADFG | 12.2 | 4.2 | 10.7 | 4.3 | 4.6 | 37.1 | 26.3 |  |
|  | Actual run | 10.4 | 5.0 | 17.6 | 5.5 | 5.2 | 45.3 | 32.0 |  |
| 1993 | FRI | 9.1 | 3.6 | 18.2 | 5.5 | 6.0 | 43.3 | 31.9 | 28 |
|  | ADFG | 11.7 | 3.4 | 15.8 | 4.9 | 5.1 | 41.8 | 32.0 |  |
|  | Actual run | 9.3 | 4.7 | 23.3 | 5.7 | 7.6 | 51.9 | 40.8 |  |

Total run and catch include Branch River and Togiak District but exclude jacks (1-ocean age).
Percent error $=$ error in forecasted catch (forecast-actual catch/forecast•100).

Table 2. Bristol Bay sockeye salmon runs and the predictions from the Port Moller test-boat catches.

| Year | Bristol Bay |  | Run pred. on $6 / 25$ |  |  | Run pred. on 6/30 |  |  | Run pred. on 7/3 |  |  | Catch pred. on 7/3 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Run | Catch | Pred. | R-P | \%ofP | Pred. | R-P | \%ofP | Pred. | R-P | \%ofP | Pred. | C-P | \%ofP |
| 1987 | 27 | 16 | 27 | 0 | 0 | 27 | 0 | 0 | 26 | 1 | 4 | 15 | 1 | 7 |
| 1988 | 23 | 14 | 15 | 8 | 53 | 15 | 8 | 53 | 22 | 1 | 5 | 12 | 2 | 17 |
| 1989 | 44 | 29 | 50 | -6 | -12 | 37 | 7 | 19 | 42 | 2 | 5 | 28 | 1 | 4 |
| 1990 | 48 | 33 | 42 | 6 | 14 | 56 | -8 | -14 | 39 | 9 | 23 | 25 | 8 | 32 |
| 1991 | 42 | 26 | 48 | -6 | -13 | 37 | 5 | 14 | 37 | 5 | 14 | 21 | 5 | 24 |
| 1992 | 45 | 32 | 49 | -4 | -8 | 45 | 0 | 0 | 41 | 4 | 10 | 29 | 3 | 10 |
| 1993 | 52 | 41 | 61 | -9 | -15 | 57 | -5 | -9 | 56 | -4 | -7 | 44 | -3 | -7 |
| Means | 40 | 27 | 42 | -2 | 3 | 39 | 1 | 9 | 38 | 3 | 7 | 25 | 2 | 12 |
| absol. |  |  |  | 6 | 16 |  | 5 | 16 |  | 4 | 10 |  | 3 | 14 |

Number in millions of fish.
$\mathrm{R}=$ run, $\mathrm{P}=$ predicted, and $\mathrm{C}=$ catch.
absol. $=$ absolute error, ignoring the sign.
$\%$ of $\mathrm{P}=$ the percentage that the actual run differed from the prediction.
1993 forecasts are from Bristol Bay almanac (not adjusted for early run timing).

Table 3. Mean lengths (mid-eye to tail fork, mm) of sockeye salmon in the Bristol Bay area.

| Year | $\begin{gathered} \text { BB run } \\ \text { (millions) } \end{gathered}$ | 2-ocean |  |  | 3-ocean |  |  | Both age groups | $\begin{aligned} & \text { Percent } \\ & \text { 3-ocean } \\ & \hline \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Male | Female | Combined | Male | Female | Combined |  |  |
| 1958 | 6 | 527 | 508 | 517 | 586 | 562 | 572 | 544 | 48 |
| 1959 | 13 | 522 | 502 | 512 | 585 | 562 | 571 | 522 | 16 |
| 1960 | 36 | 496 | 480 | 489 | 580 | 553 | 562 | 498 | 12 |
| 1961 | 18 | 525 | 512 | 519 | 583 | 562 | 572 | 554 | 66 |
| 1962 | 10 | 527 | 508 | 518 | 582 | 566 | 574 | 535 | 30 |
| 1963 | 7 | 529 | 512 | 520 | 594 | 570 | 580 | 546 | 44 |
| 1964 | 11 | 517 | 499 | 508 | 584 | 564 | 571 | 522 | 22 |
| 1965 | 53 | 506 | 487 | 497 | 574 | 552 | 561 | 502 | 8 |
| 1966 | 18 | 514 | 503 | 508 | 581 | 561 | 569 | 554 | 75 |
| 1967 | 10 | 534 | 518 | 526 | 592 | 570 | 579 | 544 | 34 |
| 1968 | 8 | 516 | 503 | 510 | 594 | 572 | 581 | 535 | 36 |
| 1969 | 18 | 524 | 510 | 517 | 591 | 571 | 580 | 525 | 22 |
| 1970 | 39 | 511 | 497 | 504 | 572 | 549 | 558 | 509 | 9 |
| 1971 | 16 | 530 | 516 | 522 | 584 | 563 | 572 | 552 | 60 |
| 1972 | 5 | 521 | 505 | 514 | 583 | 562 | 572 | 543 | 51 |
| 1973 | 2 | 522 | 513 | 518 | 601 | 575 | 587 | 575 | 82 |
| 1974 | 11 | 525 | 508 | 518 | 581 | 566 | 574 | 528 | 19 |
| 1975 | 24 | 518 | 499 | 509 | 587 | 564 | 574 | 523 | 21 |
| 1976 | 12 | 531 | 514 | 523 | 592 | 568 | 578 | 543 | 36 |
| 1977 | 10 | 533 | 517 | 525 | 597 | 573 | 584 | 556 | 53 |
| 1978 | 19 | 520 | 502 | 512 | 595 | 570 | 582 | 539 | 38 |
| 1979 | 40 | 537 | 524 | 530 | 586 | 567 | 576 | 538 | 18 |
| 1980 | 62 | 519 | 503 | 511 | 583 | 553 | 567 | 525 | 26 |
| 1981 | 34 | 536 | 523 | 529 | 588 | 566 | 577 | 555 | 54 |
| 1982 | 22 | 522 | 508 | 515 | 587 | 566 | 576 | 561 | 75 |
| 1983 | 46 | 530 | 514 | 521 | 574 | 557 | 565 | 529 | 17 |
| 1984 | 41 | 515 | 501 | 508 | 580 | 561 | 570 | 526 | 30 |
| 1985 | 37 | 527 | 512 | 520 | 583 | 567 | 575 | 543 | 41 |
| 1986 | 24 | 535 | 521 | 528 | 583 | 561 | 571 | 553 | 58 |
| 1987 | 27 | 521 | 506 | 513 | 590 | 567 | 577 | 538 | 39 |
| 1988 | 23 | 525 | 513 | 519 | 592 | 571 | 581 | 554 | 56 |
| 1989 | 44 | 525 | 507 | 515 | 586 | 564 | 575 | 538 | 27 |
| 1990 | 48 | 507 | 491 | 499 | 578 | 557 | 566 | 528 | 43 |
| 1991 | 42 | 508 | 493 | 500 | 573 | 547 | 560 | 536 | 60 |
| 1992 | 45 | 511 | 496 | 504 | 568 | 544 | 557 | 531 | 52 |
| 1993 | 52 | 530 | 515 | 522 | 582 | 560 | 570 | 547 | 52 |
| Averages |  |  |  |  |  |  |  |  |  |
| 58-67 | 18 | 520 | 503 | 511 | 584 | 562 | 571 | 532 | 36 |
| 68-77 | 15 | 523 | 508 | 516 | 588 | 566 | 576 | 539 | 39 |
| 78-87 | 35 | 526 | 511 | 519 | 585 | 564 | 574 | 541 | 40 |
| 88-92 | 45 | 513 | 497 | 505 | 576 | 553 | 565 | 533 | 46 |

Table 4. Average weights of sockeye (lbs) in the east-side Bristol Bay commercial catches, 19841993.

| District | Year $\begin{gathered}\text { Catch } \\ \text { millions }\end{gathered}$ |  | 2-ocean |  |  | 3-ocean |  |  | $\begin{gathered} \text { All } \\ \text { males } \end{gathered}$ | $\begin{gathered} \text { All } \\ \text { females } \end{gathered}$ | $\begin{array}{r} \text { All } \\ \text { fish } \\ \hline \end{array}$ | $\begin{aligned} & \text { Percent } \\ & \text { 3-ocean } \\ & \hline \end{aligned}$ | Percent females |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Male | Female | Comb. | Male | Female | Comb. |  |  |  |  |  |
| Nak/Kvi | 84 | 14.5 | 5.2 | 4.5 | 4.9 | 7.1 | 6.1 | 6.6 | 5.6 | 4.9 | 5.2 | 22 | 48 |
|  | 85 | 8.2 | 5.1 | 4.5 | 4.9 | 6.9 | 6.3 | 6.6 | 5.9 | 5.4 | 5.6 | 51 | 49 |
|  | 86 | 2.9 | 5.4 | 4.7 | 5.0 | 7.2 | 6.2 | 6.6 | 6.7 | 5.8 | 6.2 | 73 | 59 |
|  | 87 | 5.0 | 5.3 | 4.5 | 4.9 | 7.6 | 6.5 | 7.0 | 6.0 | 5.2 | 5.6 | 34 | 52 |
|  | 88 | 3.5 | 5.3 | 4.5 | 4.9 | 7.4 | 6.5 | 6.9 | 6.3 | 5.6 | 5.9 | 52 | 52 |
|  | 89 | 13.8 | 5.3 | 4.6 | 4.9 | 7.3 | 6.2 | 6.8 | 5.8 | 4.9 | 5.3 | 21 | 55 |
|  | 90 | 17.1 | 5.0 | 4.5 | 4.7 | 7.3 | 6.2 | 6.7 | 5.9 | 5.3 | 5.6 | 43 | 54 |
|  | 91 | 10.6 | 4.9 | 4.3 | 4.6 | 7.2 | 6.0 | 6.5 | 6.6 | 5.5 | 6.0 | 71 | 54 |
|  | 92 | 9.3 | 5.0 | 4.5 | 4.7 | 6.7 | 5.7 | 6.2 | 6.0 | 5.2 | 5.6 | 60 | 48 |
|  | 93 | 8.9 | 5.3 | 4.8 | 5.1 | 7.1 | 6.2 | 6.6 | 6.3 | 5.6 | 5.9 | 54 | 53 |
|  | Means | 9.5 | 5.2 | 4.5 | 4.9 | 7.2 | 6.2 | 6.6 | 6.1 | 5.3 | 5.7 | 48 | 52 |
| Egegik | 84 | 5.2 | 5.1 | 4.5 | 4.9 | 7.5 | 6.8 | 7.1 | 6.0 | 5.7 | 5.8 | 43 | 44 |
|  | 85 | 7.5 | 5.6 | 4.8 | 5.2 | 7.6 | 6.5 | 7.1 | 6.4 | 5.6 | 6.0 | 44 | 48 |
|  | 86 | 4.9 | 5.8 | 5.0 | 5.4 | 7.2 | 6.3 | 6.7 | 6.2 | 5.4 | 5.8 | 31 | 56 |
|  | 87 | 5.4 | 5.2 | 5.1 | 5.2 | 7.8 | 6.5 | 7.0 | 6.4 | 5.8 | 6.1 | 48 | 55 |
|  | 88 | 6.5 | 5.4 | 4.9 | 5.2 | 7.5 | 6.7 | 7.2 | 6.6 | 6.0 | 6.3 | 57 | 45 |
|  | 89 | 8.9 | 5.2 | 4.6 | 4.9 | 7.4 | 5.9 | 6.7 | 6.0 | 5.0 | 5.5 | 33 | 51 |
|  | 90 | 10.1 | 5.3 | 4.9 | 5.1 | 7.3 | 6.1 | 6.6 | 6.3 | 5.6 | 5.9 | 54 | 52 |
|  | 91 | 6.8 | 5.3 | 4.4 | 4.9 | 7.3 | 6.0 | 6.6 | 6.4 | 5.3 | 5.8 | 55 | 52 |
|  | 92 | 15.7 | 4.7 | 4.1 | 4.5 | 6.6 | 5.8 | 6.2 | 5.6 | 5.0 | 5.4 | 51 | 44 |
|  | 93 | 21.8 | 5.5 | 4.8 | 5.1 | 7.1 | 6.2 | 6.6 | 6.3 | 5.6 | 5.9 | 52 | 54 |
|  | Means | 9.3 | 5.3 | 4.7 | 5.0 | 7.3 | 6.3 | 6.8 | 6.2 | 5.5 | 5.9 | 47 | 50 |
| Ugashik | 84 | 2.7 | 5.0 | 4.6 | 4.9 | 6.9 | 6.2 | 6.5 | 5.6 | 5.5 | 5.6 | 42 | 35 |
|  | 85 | 6.5 | 5.6 | 4.7 | 5.2 | 7.3 | 6.3 | 6.9 | 6.2 | 5.4 | 5.8 | 38 | 43 |
|  | 86 | 5.0 | 5.9 | 5.0 | 5.5 | 7.8 | 6.4 | 7.1 | 6.9 | 5.8 | 6.2 | 55 | 49 |
|  | 87 | 2.1 | 5.5 | 4.9 | 5.2 | 7.9 | 6.7 | 7.3 | 6.9 | 6.0 | 6.5 | 61 | 47 |
|  | 88 | 1.5 | 5.4 | 4.8 | 5.2 | 7.5 | 6.6 | 7.1 | 6.4 | 5.9 | 6.2 | 54 | 43 |
|  | 89 | 3.1 | 5.5 | 4.7 | 5.1 | 7.7 | 6.5 | 7.2 | 5.9 | 5.0 | 5.5 | 19 | 45 |
|  | 90 | 2.1 | 5.0 | 4.5 | 4.7 | 7.4 | 6.4 | 6.9 | 6.1 | 5.6 | 5.9 | 53 | 49 |
|  | 91 | 3.0 | 5.3 | 4.5 | 4.9 | 7.0 | 5.8 | 6.3 | 6.2 | 5.3 | 5.8 | 59 | 52 |
|  | 92 | 3.4 | 5.0 | 4.5 | 4.8 | 6.8 | 5.6 | 6.4 | 6.2 | 5.2 | 5.8 | 64 | 37 |
|  | 93 | 4.3 | 5.7 | 4.6 | 5.2 | 7.7 | 6.7 | 7.2 | 6.7 | 5.7 | 6.2 | 52 | 52 |
|  | Means | 3.3 | 5.4 | 4.7 | 5.1 | 7.4 | 6.3 | 6.9 | 6.3 | 5.6 | 5.9 | 50 | 45 |

Table 5. Average weights of sockeye (lbs) in the west-side Bristol Bay commercial catches, 1984-1993.

| District | Year | Catch millions | 2-ocean |  |  | 3-ocean |  |  | $\begin{gathered} \text { All } \\ \text { males } \\ \hline \end{gathered}$ | $\begin{gathered} \text { All } \\ \text { females } \end{gathered}$ | $\begin{aligned} & \text { All } \\ & \text { fish } \\ & \hline \end{aligned}$ | Percent <br> 3-ocean | Percent females |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Male | Female | Comb. | Male | Female | Comb. |  |  |  |  |  |
| Nushagak | 84 | 2.0 | 5.3 | 5.1 | 5.2 | 7.4 | 6.0 | 6.6 | 7.0 | 5.9 | 6.4 | 87 | 56 |
|  | 85 | 1.3 | 5.2 | 4.6 | 4.9 | 7.4 | 6.3 | 6.8 | 6.7 | 5.8 | 6.3 | 70 | 49 |
|  | 86 | 2.7 | 4.7 | 4.5 | 4.6 | 7.3 | 6.1 | 6.6 | 6.9 | 5.9 | 6.3 | 86 | 55 |
|  | 87 | 3.3 | 5.2 | 4.5 | 4.9 | 8.3 | 6.5 | 7.2 | 6.9 | 6.0 | 6.4 | 65 | 53 |
|  | 88 | 1.7 | 4.9 | 4.3 | 4.7 | 7.8 | 6.2 | 7.0 | 7.1 | 5.9 | 6.5 | 79 | 49 |
|  | 89 | 2.8 | 5.4 | 4.3 | 4.7 | 7.6 | 6.2 | 6.8 | 6.9 | 5.6 | 6.1 | 68 | 62 |
|  | 90 | 3.6 | 4.5 | 4.1 | 4.4 | 7.6 | 5.9 | 6.7 | 6.6 | 5.5 | 6.0 | 71 | 50 |
|  | 91 | 5.3 | 4.3 | 3.8 | 4.0 | 7.1 | 5.7 | 6.3 | 6.4 | 5.2 | 5.7 | 75 | 56 |
|  | 92 | 2.8 | 4.7 | 4.0 | 4.4 | 6.5 | 5.4 | 6.0 | 5.7 | 5.0 | 5.4 | 61 | 45 |
|  | 93 | 5.3 | 5.2 | 4.3 | 4.8 | 7.5 | 6.0 | 6.6 | 6.4 | 5.4 | 5.9 | 59 | 55 |
|  | Means | 3.1 | 4.9 | 4.4 | 4.7 | 7.4 | 6.0 | 6.7 | 6.6 | 5.6 | 6.1 | 72 | 53 |
| Togiak | 84 | 0.2 | 5.7 | 5.1 | 5.4 | 7.6 | 6.0 | 6.8 | 7.4 | 6.0 | 6.6 | 89 | 54 |
|  | 85 | 0.1 | 5.0 | 4.4 | 4.6 | 7.7 | 6.0 | 6.7 | 7.3 | 5.8 | 6.4 | 85 | 59 |
|  | 86 | 0.2 | 5.8 | 4.7 | 5.2 | 7.4 | 6.0 | 6.6 | 7.1 | 5.8 | 6.4 | 84 | 55 |
|  | 87 | 0.3 | 5.9 | 4.9 | 5.5 | 8.6 | 6.9 | 7.6 | 7.5 | 6.4 | 6.9 | 68 | 55 |
|  | 88 | 0.7 | 6.3 | 5.1 | 5.6 | 8.8 | 7.2 | 7.9 | 8.7 | 7.1 | 7.8 | 97 | 54 |
|  | 89 | 0.1 | 5.9 | 4.7 | 5.4 | 8.4 | 6.3 | 7.1 | 7.8 | 6.1 | 6.8 | 82 | 57 |
|  | 90 | 0.2 | 5.4 | 4.8 | 5.0 | 8.1 | 6.3 | 7.1 | 7.7 | 6.1 | 6.8 | 85 | 57 |
|  | 91 | 0.5 | 5.9 | 4.8 | 5.4 | 8.1 | 6.2 | 7.1 | 7.4 | 5.8 | 6.6 | 69 | 50 |
|  | 92 | 0.6 | 5.4 | 4.8 | 5.1 | 8.7 | 6.3 | 7.6 | 8.2 | 6.1 | 7.2 | 85 | 47 |
|  | 93 | 0.5 | 6.2 | 5.0 | 5.6 | 9.2 | 6.5 | 7.9 | 8.5 | 6.2 | 7.3 | 76 | 49 |
|  | Means | 0.3 | 5.7 | 4.8 | 5.3 | 8.3 | 6.4 | 7.2 | 7.7 | 6.1 | 6.9 | 82 | 54 |

Table 6. Sockeye salmon escapements in excess of management goals for Bristol Bay river (in millions), 1987-1993.

| River system | Escapement goal |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Midpoint | Upper <br> range | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 |
|  |  |  | Escapement in excess of mid-point |  |  |  |  |  |  |
| Kvichak | variable | variable |  |  |  |  |  |  |  |
| Branch |  |  |  |  |  |  |  |  |  |
| Naknek | 1.00 | 1.40 | . 06 | . 04 | . 16 | 1.09 | 2.57 | . 61 | . 54 |
| Egegik | 1.00 | 1.20 | . 27 | . 61 | . 61 | 1.19 | 1.79 | . 95 | . 52 |
| Ugashik | . 70 | . 90 | . 00 | . 00 | 1.01 | . 05 | 1.76 | 1.76 | . 71 |
| Wood | 1.00 | 1.20 | . 34 | . 00 | . 19 | . 07 | . 16 | . 29 | . 18 |
| Igushik | . 20 | . 25 | . 00 | . 00 | . 26 | . 17 | . 56 | . 10 | . 21 |
| Nuyakuk/Nush. | . 50 | . 76 | . 00 | . 00 | . 01 | . 17 | . 00 | . 20 | . 21 |
| Togiak | . 15 | . 25 | . 13 | . 16 | . 00 | . 04 | . 13 | . 07 | . 04 |
| Total |  |  | . 80 | . 81 | 2.24 | 2.78 | 6.97 | 3.98 | 2.41 |
| Bristol Bay run | - |  | 27 | 23 | 44 | 48 | 42 | 45 | 52 |
| catch |  |  | 16 | 14 | 29 | 33 | 26 | 32 | 41 |
|  |  |  | Escapement in excess of upper range |  |  |  |  |  |  |
| Naknek | 1.00 | 1.40 | . 00 | . 00 | . 00 | . 69 | 2.18 | . 21 | . 14 |
| Egegik | 1.00 | 1.20 | . 07 | . 41 | . 41 | . 99 | 1.59 | . 75 | . 32 |
| Ugashik | . 70 | . 90 | . 00 | . 00 | . 81 | . 00 | 1.58 | 1.56 | . 51 |
| Wood | 1.00 | 1.20 | . 14 | . 00 | . 00 | . 00 | . 00 | . 09 | . 00 |
| Igushik | . 20 | . 25 | . 00 | . 00 | . 21 | . 12 | . 51 | . 05 | . 16 |
| Nuyakuk/Nush. | . 50 | . 76 | . 00 | . 00 | . 00 | . 00 | . 00 | . 00 | . 00 |
| Togiak | . 15 | . 25 | . 03 | . 06 | . 00 | . 00 | . 03 | . 00 | . 00 |
| Total |  |  | . 24 | .47 | 1.43 | 1.80 | 5.89 | 2.66 | 1.13 |

Table 7. Cumulative daily escapements of sockeye salmon in the Kvichak and Newhalen Rivers, 1989-1993 (numbers in 1,000s and Newhalen escapements estimated from expanded counts lagged back 2 days).

| Date | 1989 |  | 1990 |  | 1991 |  | 1992 |  | 1993 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Kvichak | Newhalen | Kvichak | Newhalen | Kvichak | Newhalen | Kvichak | Newhalen | Kvichak | Newhalen |
| 6/22 |  |  |  |  |  |  |  |  | 13 |  |
| 23 |  |  |  |  |  |  | 0 |  | 24 |  |
| 24 | 0 |  | 0 |  |  |  | 1 |  | 34 |  |
| 25 | 58 | 17 | 1 |  | 0 |  | 2 |  | 51 | 6 |
| 26 | 298 | 97 | 2 |  | 1 |  | 10 |  | 121 | 67 |
| 27 | 525 | 162 | 3 |  | 3 |  | 17 |  | 317 | 78 |
| 28 | 653 | 200 | 5 | 0 | 50 | 7 | 81 | 5 | 559 | 157 |
| 29 | 892 | 454 | 8 | 1 | 125 | 46 | 255 | 18 | 847 | 237 |
| 30 | 1509 | 641 | 39 | 2 | 277 | 95 | 446 | 67 | 932 | 394 |
| 7/1 | 2052 | 712 | 46 | 37 | 588 | 146 | 635 | 88 | 1014 | 492 |
| 2 | 2566 | 785 | 219 | 66 | 901 | 188 | 754 | 104 | 1081 | 650 |
| 3 | 3287 | 892 | 825 | 90 | 1256 | 330 | 798 | 132 | 1182 | 816 |
| 4 | 4378 | 1185 | 1412 | 110 | 1581 | 517 | 1093 | 196 | 1307 | 937 |
| 5 | 5418 | 1287 | 1874 | 139 | 1925 | 620 | 1663 | 273 | 1678 | 1022 |
| 6 | 5947 | 1358 | 2399 | 204 | 2141 | 805 | 2244 | 329 | 2372 | 1103 |
| 7 | 6611 | 1567 | 2901 | 304 | 2208 | 1132 | 2688 | 406 | 2733 | 1121 |
| 8 | 7182 | 1962 | 3509 | 375 | 2277 | 1531 | 2880 | 534 | 2932 | 1134 |
| 9 | 7518 | 2317 | 4061 | 459 | 2355 | 1721 | 2960 | 661 | 3101 | 1163 |
| 10 | 7670 | 2478 | 4692 | 648 | 2633 | 2048 | 2985 | 840 | 3264 | 1189 |
| 11 | 7708 | 2614 | 5081 | 790 | 3080 | 2202 | 3175 | 977 | 3402 | 1220 |
| 12 | 7755 | 2728 | 5388 | 961 | 3460 |  | 3662 | 1057 | 3574 | 1268 |
| 13 | 7806 | 2829 | 5803 | 1079 | 3724 |  | 4066 | 1158 | 3751 | 1322 |
| 14 | 7860 | 2944 | 6208 | 1193 | 3822 |  | 4330 | 1258 | 3818 | 1353 |
| 15 | 7914 |  | 6418 | 1297 | 3909 |  | 4438 | 1434 | 3864 |  |
| 16 | 8060 |  | 6510 |  | 3999 |  | 4517 | 1491 | 3894 |  |
| 17 | 8130 |  | 6603 |  | 4063 |  | 4578 |  | 3921 |  |
| 18 | 8164 |  | 6674 |  | 4098 |  | 4626 |  | 3958 |  |
| 19 | 8205 |  | 6733 |  | 4132 |  | 4685 |  | 3986 |  |
| 20 | 8245 |  | 6781 |  | 4166 |  | 4695 |  | 3996 |  |
| 21 | 8273 |  | 6827 |  | 4193 |  | 4710 |  | 4008 |  |
| 22 | 8287 |  | 6876 |  | 4213 |  | 4720 |  | 4016 |  |
| 23 | 8295 |  | 6915 |  | 4220 |  | 4726 |  | 4021 |  |
| 24 | 8302 |  | 6941 |  |  |  |  |  | 4024 |  |
| 25 | 8312 |  | 6970 |  |  |  |  |  | 4025 |  |

Table 8. The Kvichak lake system escapements and the percentages going to the Newhalen River and Lake Clark.

| Year | Kvichak system escapement (millions) | Newhalen/Lake Clark escapement (millions) | Percent of Kvichak (\%) | Newhalen River spawners (millions) | Lake Clark escapement (millions) | Percent of Kvichak (\%) | Tazimina River aerial count (thousands) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1979 | 11.22 | 9.00 | 80 | 0.56 | 8.44 | 75 | 504 |
| 1980 | 22.51 | 7.50 | 33 | 2.64 | 4.86 | 22 | 128 |
| 1981 | 1.75 | 0.26 | 15 | 0.03 | 0.23 | 13 | 28 |
| 1982 | 1.14 | 0.34 | 30 | 0.13 | 0.21 | 18 | 31 |
| 1983 | 3.57 | 1.08 | 30 | 0.41 | 0.67 | 19 | 212 |
| 1984 | 10.49 | 3.20 | 31 | 0.67 | 2.53 | 24 | 366 |
| 1985 | 7.21 | 1.62 | 22 | 0.15 | 1.47 | 20 | 186 |
| 1986 | 1.18 | 0.29 | 25 | 0.01 | 0.28 | 24 | 7 |
| 1987 | 6.07 | - | - | 1.46 | - | - | 246 |
| 1988 | 4.06 | 2.41 | 59 | 0.29 | 2.12 | 52 | 83 |
| 1989 | 8.32 | 2.59 | 31 | 0.10 | 2.49 | 30 | 30 |
| 1990 | 6.97 | 1.09 | 16 | 0.07 | - | - | 4 |
| 1991 | 4.22 | 1.93 | 46 | 0.10 | - | - | 16 |
| 1992 | 4.73 | 1.05 | 22 | 0.01 | 1.04 | 22 | 13 |
| 1993 | 4.03 | 1.55 | 38 | 0.01 | 1.54 | 38 | 38 |

Newhalen River spawners estimated by two times the aerial survey estimate.
Italics $=$ estimate of missing data.

Table 9. Age compositions of sockeye salmon on the Kvichak spawning grounds in 1993.

| Spawning ground | Sex | Sample size ( n ) | Age composition (\%) |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | 2.1 | 1.2 | 2.2 | 0.3 | 1.3 | 2.3 | 1.4 |
| Copper River | M | 88 | 11.4 | 26.1 | 30.7 |  | 26.1 | 5.7 |  |
|  | F | 75 |  | 14.7 | 54.6 |  | 14.7 | 16.0 |  |
| Chinkelyes Creek | M | 100 | 1.0 | 16.0 | 73.0 |  | 6.0 | 4.0 |  |
|  | F | 96 |  | 9.4 | 75.0 |  | 12.5 | 3.1 |  |
| Newhalen River | M | 93 | 2.1 | 6.4 | 36.6 | 2.1 | 40.9 | 10.8 | 1.1 |
|  | F | 78 |  | 12.8 | 24.4 |  | 48.7 | 14.1 |  |
| Tazimina River | M | 99 |  | 3.0 | 47.5 |  | 34.3 | 15.2 |  |
|  | F | 94 |  | 14.9 | 37.2 |  | 42.6 | 5.3 |  |
| Woody Is. beaches | M | 49 | 2.0 | 30.6 | 59.2 |  | 8.2 | 0.0 |  |
|  | F | 40 |  | 42.5 | 32.5 |  | 22.5 | 2.5 |  |
| Fuel Dump Is. beach | M | 52 |  | 25.0 | 48.1 |  | 23.1 | 3.8 |  |
|  | F | 50 |  | 26.0 | 60.0 |  | 12.0 | 2.0 |  |
| Knudson Bay beach | M | 94 |  | 28.7 | 36.2 |  | 30.9 | 4.2 |  |
|  | F | 108 |  | 26.9 | 40.7 |  | 28.7 | 3.7 |  |
| Pedro Creek | M | 98 | 3.1 | 54.1 | 24.5 |  | 18.4 | 0.0 |  |
|  | F | 97 |  | 42.3 | 23.7 |  | 33.0 | 1.0 | . |
| Kvichak escapement (ADF\&G,Igiugig) | M | 1630 | 5.7 | 21.1 | 41.6 | 1.4 | 24.5 | 5.2 | 0.1 |
|  | F | 1614 | 0.1 | 23.9 | 46.1 | 1.4 | 23.9 | 4.5 |  |

Kvichak escapement: also for males; $0.3 \%$ age 0.2 and $0.1 \%$ age 1.1 .

Table 10. Spawning ground estimates of sockeye salmon on 29 selected spawning grounds in Lake Iliamna and the Newhalen River system, 1956-1993.

| Year | Aerial survey counts ( $1,000 \mathrm{~s}$ ) |  |  |  |  | Tower count escapement (1,000s) | Aerial count/ escapement (\%) | Aerial <br> observer |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Beaches |  | Total |  |  |  |
|  | Rivers | Creeks | Mainland | Island |  |  |  |  |
| 56 | 775 | - | - | - |  | 9443 |  | 1 |
| 57 | 170 | - | - | - |  | 2843 |  | 1 |
| 58 | 44 | - | - | - |  | 535 |  | 1 |
| 59 | 84 | - | - | - |  | 680 |  | 1 |
| 60 | 841 | - | - | - |  | 14630 |  | 1 |
| 61 | 246 | 40 | 50 | 127 | 463 | 3706 | 12.5 | 2 |
| 62 | 140 | 52 | 21 | 12 | 225 | 2581 | 8.7 | 2 |
| 63 | 31 | 13 | 5 | 7 | 56 | 339 | 16.5 | 2 |
| 64 | 36 | 38 | 3 | 21 | 98 | 957 | 10.2 | 2 |
| 65 | 734 | 538 | 261 | 1352 | 2885 | 24326 | 11.9 | 2 |
| 66 | 248 | 153 | 134 | 46 | 581 | 3776 | 15.4 | 2 |
| 67 | 370 | 63 | 85 | 16 | 534 | 3216 | 16.6 | 3 |
| 68 | 131 | 64 | 14 | 64 | 273 | 2557 | 10.7 | 3 |
| 69 | 192 | 168 | 40 | 102 | 502 | 8394 | 6.0 | 3 |
| 70 | 790 | 574 | 216 | 506 | 2086 | 13935 | 15.0 | 3 |
| 71 | 177 | 194 | 27 | 50 | 448 | 2387 | 18.8 | 3 |
| 72 | 89 | 50 | 15 | 9 | 163 | 1010 | 16.1 | 3 |
| 73 | 35 | 18 | 6 | 6 | 65 | 227 | 28.6 | 3 |
| 74 | 294 | 269 | 72 | 122 | 757 | 4433 | 17.1 | 3 |
| 75 | 936 | 440 | 225 | 412 | 2013 | 13140 | 15.3 | 3 |
| 76 | 144 | 55 | 19 | 45 | 263 | 1965 | 13.4 | 3 |
| 77 | 124 | 20 | 88 | 28 | 260 | 1341 | 19.4 | 3 |
| 78 | 510 | 100 | 42 | 6 | 658 | 4149 | 15.9 | 3 |
| 79 | 1424 | 372 | 252 | 81 | 2129 | 11218 | 19.0 | 3 |
| 80 | 2189 | 317 | 77 | 201 | 2784 | 22505 | 12.4 | 3 |
| 81 | 187 | 85 | 16 | 20 | 308 | 1754 | 17.6 | 3 |
| 82 | 255 | 68 | 27 | 9 | 359 | 1135 | 31.6 | 3 |
| 83 | 743 | 123 | 75 | 9 | 950 | 3570 | 26.6 | 3 |
| 84 | 1902 | 359 | 597 | 84 | 2942 | 10491 | 28.0 | 4 |
| 85 | 672 | 296 | 260 | 247 | 1475 | 7211 | 20.5 | 4 |
| 86 | 57 | 16 | 12 | 5 | 90 | 1200 | 7.5 | 5 |
| 87 | 1313 | 111 | 397 | 123 | 1944 | 6100 | 31.9 | 5 |
| 88 | 481 | 123 | 116 | 15 | 735 | 4065 | 18.1 | 6 |
| 89 | 386 | 88 | 31 | 8 | 513 | 8318 | 6.2 | 6 |
| 90 | 138 | 50 | 19 | 26 | 233 | 6970 | 3.3 | 6 |
| 91 | 196 | 111 | 18 | 19 | 344 | 4223 | 8.1 | 7 |
| 92 | 198 | 151 | 35 | 19 | 403 | 4726 | 8.5 | 7 |
| 93 | 225 | 128 | 42 | 10 | 405 | 4025 | 10.1 | 7 |
| Means |  |  |  |  |  |  |  |  |
| 61-66 | 239 | 139 | 79 | 261 | 718 | 5948 | 12.5 | 2 |
| 67-83 | 505 | 175 | 76 | 99 | 856 | 5702 | 17.6 | 3 |
| 88-93 | 271 | 109 | 44 | 16 | 439 | 5388 | 9.1 | 6,7 |

Table 11. Mean townet catches (geometric means of 20 -min tows) and lengths on Sept. 1 (live, mm ) of sockeye salmon fry in lakes Iliamna and Clark.

| Brood Year | Kvichak escapement (millions) | Lake Iliamna |  | Lake Clark |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Mean catch | $\begin{array}{r} \text { Mean } \\ \text { length } \end{array}$ | Mean catch | $\begin{array}{r} \text { Mean } \\ \text { length } \end{array}$ |
| 61 | 3.7 | 90 | 53 | 13 | 50 |
| 62 | 2.6 | 12 | 45 | 54 | 50 |
| 63 | 0.3 | 5 | 54 | 3 | 50 |
| 64 | 1.0 | 7 | 62 | 2 | 50 |
| 65 | 24.3 | 170 | 53 | 23 | 52 |
| 66 | 3.8 | 67 | 57 | 15 | 47 |
| 67 | 3.2 | 78 | 62 | 47 | 59 |
| 68 | 2.6 | 43 | 62 | 9 | 50 |
| 69 | 8.4 | 386 | 61 | 11 | 55 |
| 70 | 13.9 | 127 | 44 | 20 | 38 |
| 71 | 2.4 | 4 | 50 | 15 | 41 |
| 72 | 1.0 | 3 | 58 | 17 | 48 |
| 73 | 0.2 | 2 | 71 | 12 | 57 |
| 74 | 4.4 | 491 | 54 | 80 | 55 |
| 75 | 13.1 | 252 | 49 | 105 | 49 |
| 76 | 2.0 | 16 | 53 | - | - |
| 77 | 1.3 | 11. | 61 | - | - |
| 78 | 4.1 | 339 | 62. | 65 | 56 |
| 79 | 11.2 | 282 | 53 | 60 | 48 |
| 80 | 22.5 | 134 | 61 | 26 | 59 |
| 81 | 1.8 | 37 | 52 | 58 | 46 |
| 82 | 1.1 | 9 | 68 | 18 | 57 |
| 83 | 3.6 | 242 | 64 | 40 | 56 |
| 84 | 10.5 | 147 | 46 | 84 | 51 |
| 85 | 7.2 | 63 | 54 | 16 | 49 |
| 86 | 1.2 | 10 | 60 | - | - |
| 87 | 6.1 | 79 | 63 | 11 | 56 |
| 88 | 4.1 | 22 | 58 | 21 | 48 |
| 89 | 8.3 | 181 | 55 | 19 | 47 |
| 90 | 7.0 | 336 | 54 | - | - |
| 91 | 4.2 | - | 56 | 20 | 47 |
| 92 | 4.7 | 135 | 57 | 27 | 61 |
| 93 | 4.0 |  |  |  |  |

Lake Iliamna tows in areas 7 and 8 only.

Table 12. Summary of 1993 measurements in Lake Aleknagik (Wood River Lakes).

| Measurement and first year measured | Dates | 1993 | Past years |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  |  | Average | Range |
| 1. Date of ice breakup 1949- |  | 5/1 | 5/31 | 5/14-6/16 |
| 2. Water temperature,$\begin{aligned} & 0-20 \mathrm{~m}(\mathrm{C}) \\ & 1958- \end{aligned}$ | $6 / 22$ | 8.8 | 5.8 | 3.7, 9.2 |
|  | 7/13 | 11.1 | 8.3 | 5.7, 12.0 |
|  | 8/3 | 13.1 | 10.7 | 7.7, 14.0 |
|  | $9 / 2$ | 12.3 | 11.2 | $9.3,13.0$ |
| 3. Water transparency Secchi depth (m) 1962- | $6 / 22$ | 5.4 | 8.2 | $5.3,10.5$ |
|  | 7/13 | 8.5 | 8.2 | 5.0, 10.9 |
|  | 8/3 | 7.8 | 9.3 | $6.3,11.9$ |
|  | $9 / 2$ | 7.0 | 8.7 | 5.8, 12.1 |
| 4. Water conductivity (micromhos/cm) 1968- | 6/22 | 37.8 | 38.7 | 34.7, 52.1 |
|  | 7/13 | 40.0 | 37.4 | 33.5, 42.6 |
|  | 8/3 | 38.2 | 37.1 | 32.5, 40.5 |
|  | 9/2 | 37.6 | 38.3 | 34.8, 42.5 |
| 5. Average daily solar radiation ( $\mathrm{gm} / \mathrm{cal} / \mathrm{cm}$ ) 1963- | June 1-15 | 376 | 409 | 305, 588 |
|  | June 16-30 | 397 | 410 | 265, 572 |
|  | July 1-15 | 356 | 393 | 284, 543 |
|  | July 16-31 | 406 | 354 | 194, 481 |
|  | Aug. 1-15 | 252 | 301 | 203, 402 |
|  | Aug. 15-31 | 250 | 255 | 170, 421 |
|  | Sept. 1-15 | 160 | 210 | 114, 282 |
| 6. Lake level (cm) of Lake Nerka 1952- | June 1-15 | 180 | 141 | 84, 222 |
|  | June 15-30 | 139 | 152 | 97, 218 |
|  | July 1-15 | 119 | 133 | 75, 199 |
|  | July 16-31 | 92 | 107 | 54, 172 |
|  | Aug. 1-15 | 69 | 86 | 34, 173 |
|  | Aug. 15-31 | 85 | 83 | 30, 184 |
|  | Sept. 1-15 | 129 | 83 | 29, 161 |
| $\begin{aligned} & \text { 7. Chlorophyll "a", } \\ & 0-20 \mathrm{~m}(\mathrm{mg} / \mathrm{m} 2) \\ & 1963- \end{aligned}$ | 6/22 | 34 | 30 | 10,45 |
|  | $7 / 4$ | 24 |  |  |
|  | 7/13 | 22 | 28 | 10, 43 |
|  | 7/21 | 20 |  |  |
|  | 8/3 | 18 | 23 | 6,36 |
|  | 8/12 | 24 |  |  |
|  | $8 / 21$ | 27 |  |  |
|  | $9 / 2$ | 22 | 24 | 12, 37 |
| 8. Zooplankton volume$0-60 \mathrm{~m}(\mathrm{ml} / \mathrm{m} 2)$1967- | 6/22 | 83 | 51 | 20,168 |
|  | 7/4 | 64 |  |  |
|  | 7/13 | 162 | 80 | 45-161 |
|  | 7/21 | 94 |  |  |
|  | 8/3 | 92 | 119 | 43-226 |
|  | 8/12 | 75 |  |  |
|  | 8/21 | 45 |  |  |
|  | $9 / 2$ | 56 | 63 | 26-107 |

Table 13. Five-day averages of catches of emergent midges and water temperatures at three stations on Lake Aleknagik, 1993.

| 5-day period | Catch per day |  |  |  |  |  |  | Water temperature ( ${ }^{\circ} \mathrm{C}$ ) |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1993 |  |  |  | 1969-92 |  |  | 1993 |  |  |  | 1969-92 |  |  |
|  | W | H | B | Mean | Mean | Min | Max | W | H | B | Mean | Mean | Min | Max |
| 6/1-5 | 4 | 1 | 3 | 3 |  |  |  | 8.0 | 7.0 | 11.5 | 8.8 | 2.3 | 0.0 | 9.8 |
| 6-10 | 3 | 1 | 1 | 2 | 13 | 0 | 70 | 7.5 | 7.6 | 10.4 | 8.5 | 5.0 | 0.0 | 10.4 |
| 11-15 | 1 | 5 | 1 | 2 | 13 | 1 | 53 | 7.3 | 8.1 | 10.2 | 8.5 | 6.4 | 1.0 | 9.2 |
| 16-20 | 0 | 3 | 4 | 2 | 18 | 1 | 168 | 10.1 | 9.7 | 11.3 | 10.4 | 8.0 | 3.9 | 12.7 |
| 21-25 | 4 | 3 | 17 | 8 | 7 | 0 | 42 | 12.1 | 12.8 | 13.3 | 12.7 | 8.5 | 4.8 | 12.8 |
| 26-30 | 5 | 5 | 12 | 7 | 5 | 0 | 12 | 7.1 | 10.6 | 11.6 | 9.8 | 9.9 | 6.0 | 13.9 |
| 7/1-5 | 9 | 6 | 21 | 12 | 5 | 1 | 15 | 6.8 | 9.0 | 9.8 | 8.5 | 11.1 | 7.7 | 15.5 |
| 6-10 | 13 | 10 | 58 | 27 | 9 | 2 | 24 | 11.7 | 12.5 | 13.3 | 12.5 | 11.9 | 9.7 | 15.8 |
| 11-15 | 4 | 40 | 9 | 18 | 14 | 1 | 34 | 13.9 | 14.2 | 15.8 | 14.6 | 12.2 | 9.2 | 15.9 |
| 16-20 | 8 | 30 | 18 | 19 | 14 | 2 | 36 | 16.9 | 16.3 | 16.6 | 16.6 | 12.0 | 8.5 | 17.0 |
| 21-25 | 12 | 10 | 27 | 16 | 19 | 2 | 50 | 14.7 | 16.4 | 17.0 | 16.0 | 12.6 | 7.9 | 17.2 |
| 26-30 | 16 | 3 | 10 | 10 | 30 | 8 | 58 | 15.3 | 16.1 | 16.8 | 16.1 | 13.2 | 8.9 | 15.7 |
| 31-4 | 4 | 3 | 5 | 4 | 30 | 4 | 77 | 14.5 | 16.3 | 17.0 | 15.9 | 13.5 | 10.2 | 17.5 |
| 8/5-9 | 3 | 9 | 6 | 6 | 22 | 3 | 80 | 15.9 | 16.0 | 16.7 | 16.2 | 13.5 | 10.5 | 17.1 |
| 10-14 | 1 | 3 | 1 | 2 | 16 | 2 | 54 | 16.0 | 16.4 | 16.3 | 16.2 | 13.3 | 9.5 | 18.8 |
| 15-19 | 1 | 1 | 1 | 1 | 15 | 2 | 70 | 13.5 | 13.6 | 13.9 | 13.7 | 13.2 | 11.0 | 15.7 |
| 20-24 | 0 | 1 | 0 | 0 | 7 | 1 | 28 | 12.3 | 12.0 | 12.0 | 12.1 | 13.7 | 12.0 | 15.4 |
| 25-29 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

W = Whitefish Bay; H = Hansen Bay; B = Bear Bay.

Table 14. Average catches, lengths, and growth rates for sockeye fry and age-1 threespine stickleback in Lake Aleknagik.

| Year | Sockeye salmon fry |  |  |  |  | Sockeye Escapement in year-1 (1000s) | Threespine stickleback |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean beach seine catch | Mean length on $6 / 23$ $(\mathrm{~mm})$ | $\begin{gathered} \text { Mean } \\ \text { length } \\ \text { on } 9 / 1 \\ (\mathrm{~mm}) \end{gathered}$ | $\begin{array}{r} \hline \text { Growth } \\ \text { rate } \\ \text { (mm/ } \\ \text { day) } \\ \hline \end{array}$ | $\begin{array}{r} \text { Mean } \\ \text { tow } \\ \text { net } \\ \text { catch } \end{array}$ |  | Mean beach seine catch | $\begin{array}{r} \text { Mean } \\ \text { length } \\ \text { on } 6 / 23 \\ (\mathrm{~mm}) \\ \hline \end{array}$ | $\begin{gathered} \text { Mean } \\ \text { length } \\ \text { on } 9 / 1 \\ (\mathrm{~mm}) \\ \hline \end{gathered}$ | $\begin{array}{r} \text { Growth } \\ \text { rate } \\ (\mathrm{mm} / \\ \text { day }) \\ \hline \end{array}$ | $\begin{array}{r} \text { Mean } \\ \text { tow } \\ \text { net } \\ \text { catch } \end{array}$ | Age 0 tow net catch |
| 58 | - | - | 62.1 | - | 14 | 88 | - |  | 44.6 | - | 36 | $<1$ |
| 59 | - | - | 62.7 | - | 13 | 63 | - |  | 46.7 | - | 136 | 10 |
| 60 | - | - | 55.5 | - | 111 | 205 |  |  | 43.4 |  | 53 | 2 |
| 61 | - | - | 58.4 | - | 103 | 85 |  |  | 42.0 |  | 38 | <1 |
| 62 | 334 | 31.7 | 54.1 | . 31 | 54 | 153 | 317 | 31.0 | 43.5 | . 17 | 139 | 5 |
| 63 |  | - | 62.1 |  | 24 | 48 |  |  | 46.4 |  | 46 | 1 |
| 64 | 227 | 31.1 | 60.4 | . 42 | 24 | 31 | 352 | 31.2 | 43.1 | . 17 | 272 | 1 |
| 65 | 549 | 31.2 | 53.6 | . 32 | 103 | 155 | 202 | 29.1 | 39.5 | . 15 | 182 | 1 |
| 66 | 395 | 30.2 | 47.5 | . 25 | 219 | 220 | 258 | 27.1 | 39.4 | . 18 | 150 | 0 |
| 67 | 339 | 30.7 | 43.4 | . 18 | 49 | 287 | 426 | 28.2 | 41.3 | . 19 | 61 | 5 |
| 68 | 46 | 31.8 | 57.9 | . 37 | 10 | 92 | 212 | 30.8 | 43.4 | . 18 | 268 | 169 |
| 69 | 96 | 31.7 | 61.4 | . 43 | 78 | 177 | 215 | 33.4 | 44.2 | . 16 | 81 | <1 |
| 70 | 164 | 31.4 | 59.0 | . 40 | 43 | 160 | 156 | 32.1 | 44.8 | . 18 | 87 | <1 |
| 71 | 408 | 30.6 | 54.6 | . 35 | 17 | 302 | 261 | 29.6 | 43.4 | . 20 | 3 | <1 |
| 72 | 126 | 30.6 | 54.8 | . 35 | 10 | 182 | 45 | 28.0 | 44.4 | . 24 | 12 | 1 |
| 73 | 30 | 29.0 | 66.7 | . 54 | 3 | 98 | 62 | 29.3 | 49.5 | . 29 | 8 | 1 |
| 74 | 47 | 35.3 | 62.8 | . 39 | 44 | 162 | 125 | 33.1 | 50.1 | . 24 | 119 | <1 |
| 75 | 111 | 29.1 | 55.3 | . 39 | 8 | 242 | 69 | 32.5 | 42.4 | . 15 | 132 | $<1$ |
| 76 | 178 | 30.1 | 49.8 | . 29 | 394 | 457 | 279 | 27.7 | 39.6 | . 17 | 30 | <1 |
| 77 | 223 | 30.1 | 48.0 | . 27 | 25 | 314 | 184 | 29.3 | 40.8 | . 17 | 36 | <1 |
| 78 | 34 | 32.8 | 62.7 | . 43 | 6 | 152 | 64 | 31.7 | 47.5 | . 23 | 21 | 1 |
| 79 | 312 | 31.6 | 51.5 | . 28 | 130 | 612 | 82 | 33.2 | 42.3 | . 13 | 50 | 18 |
| 80 | 46 | 31.0 | 56.4 | . 35 | 3 | 354 | 32 | 31.0 | 44.9 | . 19 | 24 | <1 |
| 81 | 423 | 32.4 | 51.3 | . 27 | 6 | 1230 | 217 | 34.7 | 45.5 | . 15 | 12 | <1 |
| 82 | 53 | 30.0 | 52.2 | . 33 | 131 | 454 | 63 | 30.2 | 43.2 | . 19 | 12 | 0 |
| 83 | 43 | 32.1 | 63.9 | . 45 | 22 | 337 | 12 | 30.9 | 48.4 | . 25 | 64 | 12 |
| 84 | 16 | 36.2 | 64.2 | . 41 | 3 | 245 | 54 | 35.9 | 48.8 | . 19 | 200 | 155 |
| 85 | 102 | 31.0 | 56.3 | . 36 | 1 | 329 | 109 | 34.3 | 40.9 | . 09 | 2 | 0 |
| 86 | 32 | 32.2 | 58.4 | . 37 | 10 | 189 | 24 | 31.4 | 45.0 | . 19 | 11 | 0 |
| 87 | 69 | 29.7 | 57.5 | . 40 | 3 | 343 | 27 | 31.7 | 44.9 | . 19 | 67 | <1 |
| 88 | 31 | 31.2 | 58.8 | . 40 | 2 | 362 | 42 | 32.4 | 48.5 | . 23 | 8 | 1 |
| 89 | 45 | 31.4 | 55.4 | . 34 | 18 | 286 | 26 | 32.6 | 47.0 | . 21 | 17 | 1 |
| 90 | 100 | 32.7 | 57.7 | . 36 | 20 | 474 | 129 | 31.2 | 48.1 | . 24 | 27 | 1 |
| 91 | 63 | 30.1 | 52.9 | . 33 | 14 | 460 | 108 | 31.3 | 42.2 | . 16 | 41 | 1 |
| 92 | 242 | 30.0 | 46.1 | . 24 | 52 | 794 | 200 | 27.9 | 39.4 | . 17 | 222 | <1 |
| 93 | 23 | 33.7 | 56.4 | . 33 | 12 | 343 | 55 | 31.7 | 46.5 | . 22 | 3 | <1 |
| Means | 158 | 31.4 | 56.0 | . 35 | 48 | 314 | 142 | 31.1 | 44.3 | . 19 | 75 | 11 |

1. Beach seine catches at 10 stations for four dates during June 22-July 14
2. Townet catches for $5-\mathrm{min}$ hauls, two at each of six stations during Sept. 1-5.
3. Lengths measured to nearest mm on preserved fish, means adjusted to live measurement.
4. Threespine stickleback catches are for all ages (0-4) but mean length for age 1 only.

Table 15. Average townet catches and mean lengths of sockeye fry (by lake area), numbers of parent spawners, and average catches and mean lengths (age 1) of threespine stickleback for Lake Nerka.

| Year | Sockeye salmon fry |  |  |  |  |  | Sockeye salmon spawners in year-1 (1000s) |  |  | Threespine stickleback |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean tow-net catch |  |  | Mean length (mm) on 9/1 |  |  |  |  |  | Mean tow- | Mean length |
|  | South | Central | North | South | Central | North | South | Central | North | net catch | (mm) on 9/1 |
| 58 | 4 | 4 | 10 | 62 | 60 | 61 | 73 | 57 | 52 | 26 | 44 |
| 59 | 17 | 9 | 4 | 66 | 61 | 61 | 163 | 58 | 188 | 35 | 43 |
| 60 | 62 | 42 | 42 | 58 | 55 | 51 | 564 | 332 | 395 | 11 | 42 |
| 61 | 108 | 57 | 64 | 59 | 56 | 54 | 231 | 137 | 214 | 8 | 41 |
| 62 | 2 | 7 | 26 | 64 | 59 | 59 | 49 | 50 | 143 | 6 | 47 |
| 63 | 58 | 18 | 55 | 62 | 60 | 62 | 97 | 73 | 126 | 9 | 48 |
| 64 | 3 | 7 | 44 | 57 | 55 | 64 | 56 | 65 | 110 | 8 | 45 |
| 65 | 15 | 8 | 93 | 57 | 54 | 54 | 110 | 159 | 161 | 9 | 40 |
| 66 | 4 | 7 | 70 | 57 | 54 | 54 | 60 | 77 | 184 | 6 | 44 |
| 67 | 8 | 18 | 58 | 64 | 58 | 59 | 149 | 141 | 246 | 12 | 46 |
| 68 | 4 | 11 | 8 | 68 | 64 | 65 | 44 | 64 | 114 | 25 | 48 |
| 69 | 15 | 4 | 27 | 65 | 61 | 60 | 46 | 103 | 150 | 14 | 46 |
| 70 | 2 | 5 | 21 | 64 | 65 | 63 | 51 | 56 | 266 | 5 | 43 |
| 71 | 3 | 9 | 197 | 54 | 52 | 58 | 141 | 132 | 229 | 4 | 42 |
| 72 | 2 | 11 | 8 | 57 | 55 | 55 | 68 | 73 | 178 | 8 | 45 |
| 73 | 1 | 3 | 11 | 61 | 61 | 61 | 37 | 82 | 109 | 4 | 45 |
| 74 | 5 | 4 | 34 | 69 | 64 | 64 | 19 | 29 | 83 | 107 | 50 |
| 75 | 7 | 15 | 9 | 59 | 55 | 53 | 236 | 141 | 242 | 60 | 44 |
| 76 | 1 | 9 | 40 | 52 | 49 | 45 | 128 | 69 | 297 | 17 | 40 |
| 77 | 19 | 50 | 143 | 55 | 54 | 51 | 77 | 69 | 176 | 17 | 42 |
| 78 | <1 | <1 | 4 | 56 | 61 | 63 | 67 | 65 | 173 | 18 | 46 |
| 79 | 3 | 17 | 50 | 64 | 54 | 58 | 151 | 181 | 460 | 61 | 47 |
| 80 | 1 | 14 | 37 | 52 | 49 | 47 | 246 | 142 | 287 | 33 | 41 |
| 81 | 3 | 16 | 13 | 59 | 55 | 55 | 219 | 224 | 566 | 6 | 46 |
| 82 | 1 | 6 | 38 | 54 | 56 | 54 | 89 | 169 | 348 | 24 | 45 |
| 83 | 2 | 4 | 4 | 66 | 63 | 63 | 29 | 43 | 396 | 1 | 48 |
| 84 | 1 | 11 | 2 | 72 | 61 | 63 | 67 | 79 | 247 | 14 | 50 |
| 85 | 1 | 2 | 123 | 61 | 56 | 55 | 62 | 84 | 377 | 2 | 45 |
| 86 | 2 | 16 | 12 | 50 | 54 | 64 | 51 | 106 | 492 | 2 | 42 |
| 87 | 1 | 7 | 21 | 57 | 56 | 55 | 35 | 65 | 253 | 4 | 43 |
| 88 | $<1$ | 2 | 7 | 64 | 57 | 57 | 77 | 213 | 293 | 2 | 49 |
| 89 | 1 | 3 | 16 | 57 | 51. | 59 | 56 | 173 | 178 | 5 | 48 |
| 90 | 1 | 7 | 3 | 63 | 62 | 58 | 87 | 154 | 380 | 3 | 48 |
| 91 | 27 | 22 | 32 | 61 | 57 | 56 | 68 | 117 | 214 | 27 | 44 |
| 92 | 4 | 16 | 10 | 57 | 55 | 55 | 52 | 44 | 99 | 4 | 41 |
| 93 | 8 | 6 | 16 | 62 | 57 | 55 | 213 | 261 | 205 | 15 | 45 |
| Means | 12 | 13 | 38 | 60 | 57 | 58 | 107 | 109 | 241 | 17 | 45 |

Table 16. Occurrence and numbers of juvenile sockeye in stomachs of Arctic char collected by hook and line from Little Togiak River during 30 days after ice-out.

| Year | Date of ice out | Range in sampling dates | Number of char examined | $\begin{aligned} & \text { Mean } \\ & \text { length } \\ & \text { (mm) } \end{aligned}$ | Percent of char with: |  | Mean number per char |  | Sockeye escape. year-2 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  | Fry | Smolt | Fry | Smolt |  |
| 72 | 6/17 | 6/26-7/10 | 82 | 446 | 34 | 60 | 2.8 | 4.5 | 55 |
| 73 | 6/08 | 6/19-7/03 | 121 | 446 | 34 | 44 | 1.9 | 2.9 | 24 |
| 74 | 5/27 | 6/11-25 | 64 | 429 | 19 | 39 | 0.8 | 1.6 | 14 |
| 75 | 6/15 | 6/22-7/13 | 71 | 415 | 9 | 36 | 0.2 | 1.8 | 14 |
| 76 | 6/17 | 6/19-7/13 | 96 | 418 | 11 | 56 | 0.4 | 2.2 | 48 |
| 77 | 6/13 | 6/11-7/11 | 325 | 403 | 30 | 17 | 7.0 | 0.4 | 30 |
| 78 | 6/02 | 6/07-25 | 316 | 437 | 7 | 42 | 0.2 | 1.5 | 18 |
| 79 | 5/24 | 6/06-22 | 178 | 438 | 32 | 25 | 1.8 | 1.2 | 26 |
| 80 | 5/27 | 6/09-25 | 278 | 459 | 27 | 81 | 1.4 | 9.4 | 45 |
| 81 | 5/28 | 6/12-25 | 124 | 415 | 3 | 31 | 0.1 | 1.4 | 44 |
| 82 | 6/15 | 6/17-7/05 | 105 | 450 | 18 | 61 | 1.8 | 6.4 | 81 |
| 83 | 5/27 | 6/19-7/03 | 78 | 424 | 0 | 14 | 0.0 | 0.3 | 60 |
| 84 | 5/26 | 6/20-7/02 | 56 | 408 | 0 | 18 | 0.0 | 0.4 | 36 |
| 85 | 6/17 | 6/15-7/06 | 60 | 437 | 22 | 30 | 1.6 | 1.2 | 31 |
| 86 | 6/04 | 6/16-7/05 | 61 | 437 | 21 | 56 | 0.4 | 2.7 | 17 |
| 87 | 6/01 | 6/14-7/05 | 51 | 451 | 6 | 78 | 0.1 | 4.9 | 19 |
| 88 | 6/05 | 6/16-29 | 43 | 431 | 7 | 26 | 0.1 | 0.8 | 18 |
| 89 | 6/17 | 6/20-7/15 | 105 | 388 | 37 | 38 | 2.2 | 1.3 | 13 |
| 90 | 5/28 | 6/07-24 | 72 | 391 | 35 | 11 | 1.8 | 0.3 | 15 |
| 91 | 6/07 | 6/20-7/07 | 48 | 415 | 4 | 35 | 0.9 | 3.2 | 11 |
| 92 | 6/13 | 6/15-7/11 | 79 | 425 | 0 | 46 | 0.0 | 1.9 | 31 |
| 93 | 5/12 | 6/07-18 | 51 | 428 | 21 | 22 | 1.4 | 0.7 | 6 |
| $\begin{gathered} 72-92 \\ \text { means } \end{gathered}$ | 6/06 |  | 115 | 427 | 17 | 40 | 1.2 | 2.4 | 31 |

Table 17. Ground survey counts of sockeye spawners in the Wood River Lakes, 1993.

| Location | Date | Estimated off mouth | In creek |  |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Live | Dead | Natural | Bear kill |  |
| Aleknagik |  |  |  |  |  |  |  |
| Yako | 8/02 | 300 | 1506 | 523 | 102 | 421 | 2329 |
| Hansen | $8 / 06$ | 200 | 1482 | 1573 | 683 | 890 | 3255 |
| Bear | 8/06 | 500 | 2402 | 1081 | 634 | 447 | 3983 |
| Happy | $8 / 07$ | 200 | 2521 | 16269 | 13217 | 3052 | 18990 |
| Ice* | 8/09 | 150 | 3449 | 2481 | 1823 | 658 | 6080 |
| Eagle | 8/10 | 200 | 1200 | 50 | 5 | 45 | 1450 |
| Mission | 8/13 | 5 | 2086 | 636 | 520 | 116 | 2727 |
| Whitefish | 8/14 | 0 | 448 | 164 | 23 | 141 | 612 |
| Nerka |  |  |  |  |  |  |  |
| Fenino | 8/11 | 10 | 408 | 4722 | 3027 | 1695 | 5140 |
| Lynx | 8/19 | 0 | 3019 | 359 | 293 | 66 | 3378 |
| Stoval1* | 8/21 | 0 | 841 | 286 | 267 | 19 | 1127 |
| Pick | 8/13 | 0 | 2962 | 2201 | 1133 | 1068 | 5163 |
| Kema* | 8/26 | 0 | no c | s, but not | eavy and | 90\% dead |  |
| Hidden Lake | 8/20 | 20 | 279 | 348 | 248 | 100 | 647 |
| Elva | 8/27 | 0 | 19 | 8 | 4 | 4 | 27 |
| Beverley |  |  |  |  |  |  |  |
| Moose* | 8/12 | 0 | 546 | 406 | 268 | 138 | 952 |
| Kulik |  |  |  |  |  |  |  |
| Grant River* | 8/24 |  | no co | but me | m densit | 95\% dead |  |

[^0]Table 18. Daily counts of sockeye salmon spawners in Hansen Creek, 1993.

| Date | $\begin{gathered} \text { Estimate } \\ \text { off } \\ \text { mouth } \\ \hline \end{gathered}$ | In creek |  |  | In ponds |  |  | Cumulative dead | $\begin{array}{r} \hline \text { Live+ } \\ \text { cum. } \\ \text { dead } \\ \hline \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Live | $\begin{array}{r} \text { Natural } \\ \text { dead } \end{array}$ | $\begin{aligned} & \text { Bear } \\ & \text { dead } \end{aligned}$ | Live | $\begin{array}{r} \text { Natural } \\ \text { dead } \end{array}$ | $\begin{aligned} & \text { Bear } \\ & \text { dead } \end{aligned}$ |  |  |
| 7/18 |  | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 1 |
| 19 |  | 0 | 1 | 1 | 0 | 0 | 0 | 3 | 3 |
| 20 |  | 303 | 6 | 3 | 0 | 0 | 0 | 12 | 315 |
| 21 |  | 216 | 3 | 28 | 0 | 0 | 0 | 43 | 259 |
| 22 |  | 148 | 2 | 19 | 2 | 0 | 0 | 64 | 214 |
| 23 |  | 143 | 3 | 11 | 0 | 0 | 0 | 78 | 221 |
| 24 |  | 69 | 2 | 53 | 0 | 0 | 0 | 133 | 202 |
| 25 |  | 13 | 7 | 36 | 0 | 0 | 0 | 176 | 189 |
| 26 |  | 512 | 15 | 16 | 0 | 0 | 0 | 207 | 719 |
| 27 |  | 427 | 7 | 38 | 10 | 0 | 0 | 252 | 689 |
| 28 |  | 947 | 11 | 57 | 47 | 0 | 0 | 320 | 1314 |
| 29 |  | 590 | 14 | 101 | 35 | 0 | 0 | 435 | 1060 |
| 30 |  | 1129 | 27 | 35 | 80 | 0 | 0 | 497 | 1706 |
| 31 |  | 1241 | 19 | 26 | 70 | 0 | 0 | 542 | 1853 |
| 8/1 |  | 1259 | 31 | 112 | 100 | 0 | 0 | 685 | 2044 |
| 2 |  | 1497 | 40 | 38 | 120 | 0 | 0 | 763 | 2380 |
| 3 |  | 1572 | 59 | 117 | 100 | 0 | 0 | 939 | 2611 |
| 4 | 100 | 1343 | 52 | 50 | 140 | 0 | 1 | 1042 | 2525 |
| 5 | 250 | 1707 | 91 | 52 | 150 | 1 | 0 | 1186 | 3043 |
| 6 | 200 | 1295 | 290 | 95 | 187 | 2 | 0 | 1573 | 3055 |
| 7 | 120 | 1133 | 121 | 94 | 200 | 0 | 0 | 1788 | 3121 |
| 8 |  | 1211 | 190 | 16 | 170 | 0 | 0 | 1994 | 3375 |
| 9 |  | 966 | 233 | 60 | 200 | 18 | 0 | 2305 | 3471 |
| 10 |  | 928 | 228 | 41 | 180 | 29 | 3 | 2606 | 3714 |
| 11 | 100 | 698 | 140 | 28 | 180 | 26 | 0 | 2800 | 3678 |
| 12 | 100 | 743 | 133 | 11 | 180 | 34 | 1 | 2979 | 3902 |
| 13 | 50 | 578 | 136 | 64 | 164 | 43 | 2 | 3224 | 3966 |
| 14 | 50 | 352 | 173 | 75 | 168 | 33 | 2 | 3507 | 4027 |
| 15 | 25 | 215 |  |  | 87 |  |  | 3647 | 3949 |
| 16 |  | 133 |  |  | 88 |  |  | 3787 | 4008 |
| 17 |  | 51 | 232 | 96 | 90 | 88 | 4 | 3927 | 4068 |
| 18 | 0 | 38 | 31 | 21 | 55 | 27 | 0 | 4006 | 4099 |
| 19 |  | 30 |  |  | 48 |  |  | 4050 | 4128 |
| 20 |  | 23 |  |  | 40 |  |  | 4093 | 4156 |
| 21 |  | 15 |  |  | 32 |  |  | 4136 | 4183 |
| 22 | 0 | 8 | 43 | 93 | 25 | 34 | 3 | 4179 | 4212 |
| Totals |  |  | 2340 | 1488 |  | 335 | 16 |  |  |

Dead fish removed on each survey.
Italics for estimates (no survey).

Table 19. Summary of Hansen Creek spawning surveys, 1990-93.

| Year | Date first fish entered | Survey date | Survey counts |  |  |  | $\begin{gathered} \text { Total } \\ \text { from } \\ \text { daily } \\ \text { surveys } \\ \hline \end{gathered}$ | Percent peak count of total | Mortalities |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  | Natural |  | $\begin{array}{r} \text { Bear- } \\ \text { kill } \end{array}$ | Percent bear- |
|  |  |  | Mouth | Live | Dead | Total |  |  | dead | dead | kill |
| 1990 | 7/28 | 8/1 | ?? | 3570 | 201 | 3771 |  | 6733 | 56 |  |  |  |
|  |  | 8/6 | 25 | 4105 | 743 | 4873 | 6733 | 72 | 5139 | 1594 | 24 |
| 1991 | 7/21 | 8/1 | ?? | 4460 | 1664 | 6124 | 16296 | 38 |  |  |  |
|  |  | $8 / 6$ | 500 | 8670 | 3735 | 12905 | 16296 | 79 | 13671 | 2625 | 16 |
| 1992 | 7/18 | 8/1 | ?? | 4594 | 1085 | 5679 | 7292 | 78 |  |  |  |
|  |  | 8/6 | 50 | 3518 | 2886 | 6454 | 7292 | 89 | 5991 | 1301 | 18 |
| 1993 | 7/20 | 8/1 | ?? | 1359 | 685 | 2044 | 4212 | 49 |  |  |  |
|  |  | 8/6 | 200 | 1482 | 1573 | 3055 | 4212 | 73 | 2675 | 1504 | 36 |


[^0]:    *Partial count; entire stream not surveyed.

